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Patterns of biological invasions in French freshwater systems by non-indigenous macroinvertebrates

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Abstract

Freshwater biodiversity is threatened by several mechanisms, of which the introduction of non-indigenous species and habitat alteration are the two most important. Exotic species act at various levels of organisation of macroinvertebrate communities, and are involved in different processes mediating their impacts on biodiversity, such as habitat modification or negative interactions with autochthonous fauna. The present work gives a list of the 43 French freshwater non-indigenous species, which represent 1.2% of the French freshwater macroinvertebrates. We provide their geographic origins, their distributions among zoological units by comparison with the native fauna and their functional characteristics according to a recent typology based on bio/ecological traits. An exponential trend of the cumulated number of non-indigenous species was evidenced, with a clumping of invaders within crustaceans and molluscs. Donor areas of non-indigenous species are in majority European, and the Ponto-Caspian basin is identified as the principal one. This pattern could be explained by a spread along waterways but its origin lies in a process of recolonisation of defaunated areas following several episodes of glaciation / deglaciation in Western Europe during the last 80,000 years. Finally, from a functional point of view, non-indigenous species exhibit a limited diversity, with two functional groups representing 80% of them.

Introduction

Biodiversity is not a static dimension that could be defined once and for all, but is, on the contrary, permanently changing, influenced by natural and anthropic processes. Among this last category, biological invasions are now considered to be a major driver of change of freshwater biodiversity (Sala *et al.*, 2000).

Invasions consist of the transfer, establishment and spread of a species within an ecosystem where it is not naturally present, also called the recipient ecosystem (Vermeij, 1996; Williamson, 1996; Kolar & Lodge, 2001). In a strict sense, the introduction of non-indigenous species occurs naturally, and is not exclusively a human-driven phenomenon (Mack *et al.*, 2000). However, a growing interest is accorded to the human-mediated invasions, since increasing rates of species establishment have been observed these last decades in many

aquatic ecosystems (The Great Lakes, Ricciardi, 2001; San-Francisco Bay, Cohen & Carlton, 1998) as a consequence of the removal of natural geographic barriers and the expanding transport by waterways. The disappearance of autochthonous species in human-altered freshwater ecosystems, the poverty of the fauna remaining and their inability to adapt to new local conditions (Sax and Brown 2000) have favored the permeability of to repeated introductions of exotic species (decrease in biotic resistance as defined by Elton 1958). Many impacts of invaders have been referenced, and among them the threat for biodiversity they represent (Sala *et al.*, 2000; Mooney & Cleland, 2001; Rahel, 2002). Effectively, invasive species can have major ecological consequences on faunal composition, community structure and ecosystem functioning (Mack *et al.*, 2000). All these impacts are mediated by numerous processes that act at various levels of organisation, from the individual to the community as a whole (Simon & Townsend, 2003).

In many countries, biological invasions are heavily studied, with national databases on non-indigenous species present, their dynamics, distributions and impacts (Leppäkoski *et al.*, 2002). When performed for freshwaters at a large scale (bij de Vaate *et al.*, 2002; Leppäkoski *et al.*, 2002), such studies do not take into account French territory. However, thanks to the high diversity of French ecoregions, non-indigenous species already referenced in southern, eastern and northern Europe could potentially be found there. Moreover, with regards to its geographic position and its high connectivity with main European waterways and other continents due to the presence of many harbours (Bordeaux, Le Havre, Marseille-Fos, Nantes-Saint-Nazaire, Rouen), France could constitute an important platform of fauna exchange. The territory includes six main basins, corresponding to the largest rivers: Garonne, Loire, Seine, Rhone, Meuse and the upper Rhine. Several streams along the Atlantic coast directly flow to the ocean and are not connected to one of the main rivers. Several rivers (Rhone, Meuse, Rhine) flow across several countries.

The present study aims to make an inventory of non-indigenous species observed in French aquatic ecosystems and to analyse their geographic origin, their distribution among freshwater fauna and their functional position according to a typology based on bio/ecological traits (Usseglio-Polatera *et al.*, 2001). These results are discussed in terms of non-indigenous species origins and potential consequences of introductions.

Material and methods

Based on a huge bibliographic analysis, we documented the non-indigenous species that could be found in French freshwater systems. Only free-living macroinvertebrates were considered in this study, *i.e.* parasites and invertebrates below 3-5 mm in their adult size (microinvertebrates as defined by Tachet *et al.*, 2000) were excluded. Data on non-indigenous species are quite difficult to collect based on the bibliography. French national agencies perform bioassessment using a global method that requires identification to the family level only, thus are unable to provide such data. Moreover, very few articles deal with species introductions in French aquatic ecosystems, and when available, such information is often published in local society of natural history bulletins, which are difficult to obtain.

Thus we kept only data signalling a new species if at least one scientific article is available (*i.e.* referenced in an international review or in a national review included in PASCAL database). For each species, we defined the date of the first observation, its origin and its major ecological and biological characteristics. Species distribution among the major French basins is estimated with bibliography, but it does not always reflect the true distribution. Consequently, it will not be possible to describe the dynamics of colonisation and dispersal of non-indigenous species within French aquatic ecosystems. For most of the species, vector of introduction is not known, even if numerous studies have already detailed

the major role of shipping traffic, and particularly ballast water transport, in the spread of aquatic exotic species throughout the world (see Gollash *et al.*, 2002 for a review).

The functional classification recently developed for benthic macroinvertebrates based on both biological and ecological traits (Usseglio-Polatera *et al.*, 2001) was used to assess the main characteristics of non-indigenous species. This classification has led to the definition of 6 functional groups, 9 subgroups and 43 units according to a biological versus ecological ordination. For species not referenced before, P. Usseglio-Polatera performed a new analysis that allowed us to integrate and classify them. The present study is based on the subgroup level and French non-indigenous species belong to 6 of them (called β , $\delta 1$, $\delta 2$, $\gamma 1$, $\gamma 2$ and ζ by Usseglio-Polatera *et al.*, 2001). Most species of the group β exhibit small to medium sizes, an omnivorous regime and a scraper or shredder feeding behaviour. They live in coarse mineral substrates and are oligo to β -mesosaprobic. Groups $\delta 1$ and $\delta 2$ shared many characteristics, such as living on organic substrates and being β -mesosaprobic. They tend to be predator, feeding on macroinvertebrates, and exhibit small to medium sizes. Species of the groups $\gamma 1$ and $\gamma 2$ live on mineral or organic substrates and are β -mesosaprobic. The feeding habits of these medium-sized taxa are similar to that of group β . Finally, group ζ is constituted of filter- or deposit-feeders that fed on microphytes and fines detritus, with variable body sizes and microhabitat preferences. All taxa of these 6 functional groups are eurythermous.

Finally, the proportion of non-indigenous species among the different zoological units that could be found in French rivers was analysed in order to assess whether some taxa have greater roles in biological invasions than others. Nematomorpha were excluded, because we did not get enough data to reliably estimate the proportion of non-indigenous species within this group. All species occurring in surface freshwaters were taken into account for this analysis.

Results

A major phenomenon

As a whole, 43 non-indigenous species have been recorded within French aquatic ecosystems. It is difficult to identify the first species that invaded these hydrosystems, because the knowledge and the study of freshwater macroinvertebrates is not an old discipline. However, the history of species establishment can be made for the last 175 years, beginning with the description of *Atyaephyra desmaresti* in the Loire basin in 1832 (Dhur & Massard, 1995), and *Gammarus roeseli* in the vicinity of Paris in 1835 (Jazdzewski, 1980) (Table 1). Since these dates, new species in French freshwater fauna have been identified regularly. The actual trend of invasion, if we consider the whole territory, could be described by a highly significant exponential function (Fig. 1), suggesting an acceleration of the phenomenon in the last few decades.

Geographic origin of non-indigenous species and distribution

Species introductions are strongly dominated by intra-continental transfers (25 of 43 species are of European origin, Table 1). Among the three European donor areas identified, we notice that the Ponto-Caspian basin distinguishes itself (32.5% of whole non-indigenous species, 58% of those with a European origin), and confirms its status as a hot-spot donor area of non-indigenous species. The two others main donor areas are North America and South-Eastern Asia. It is noteworthy that, excepting *Potamopyrgus antipodarum*, all non-indigenous species came from the northern hemisphere.

The description of the present distribution of non-indigenous species in France represents a hard task. As signalled in the introduction, few data are available for each basin.

The Rhine basin is better studied than the others, and exotic species arrivals are particularly well documented there, due both to its position among Western Europe waterways and numerous programs of rehabilitation. The Rhone basin is second in terms of available data concerning macroinvertebrates communities. It is harder to get detailed information for basins such as that of the Loire and Garonne rivers.

A strong dissymmetry between taxa

To date, 1.2 % of French freshwater macroinvertebrates is of exotic origin considering species introduced in the past 200 years. Those non-indigenous species could be found among almost all taxonomic groups. However, some of them are more implicated in biological invasions than others (Table 2). Considering the species number, two groups distinguish themselves: molluscs and crustaceans, that represent respectively 26% and 49% of all non-indigenous species found in French hydrosystems.

We also notice that non-indigenous species are represented among all zoological units, except Porifera, a weakly diversified group. The most diverse group of freshwater invertebrates, the insects, is the one where the proportion of exotics is the lowest. Conversely, freshwater Polychaeta are only represented by the non-indigenous taxon *Hypania invalida*. Molluscs and Crustaceans excepted, it seems that no taxonomical clumping of non-indigenous species occurred.

Finally, for 36 species only the functional group is known. The large majority of non-indigenous species belong to two bio/ecological groups: γ_2 and ζ (respectively 9 and 19 species in each group). Four other functional groups are represented, but by a lower number of species (2 species in each group), and, except for group β , by species that do not have spectacular population explosions within French freshwater systems (see Table 1 for species and references).

Discussion

All freshwater fauna of France could be considered as non-indigenous, going back in time enough, to the last glaciations. During the Würm glaciation (80,000-10,000 BP), the major part of Western Europe was covered with ice. Thus, at the end of this period, defaunate freshwater ecosystems appeared, colonised or recolonised by species having survived in refuge areas that were outside ice range extension (according to Oberdorff *et al.*, 1997). We can not define whether a recolonisation or a colonisation process occurred (i.e. did the species were the same before and after the glaciations or not?), because the fauna present before the Würm glaciation is poorly known (few invertebrate fossil records). The two main refuge areas in Europe are Mediterranean Europe and the Danube basin (Persat & Keith, 1997). The process of colonisation, which is well documented for fish, probably also took place for macroinvertebrates.

Thus, the dominant pattern of invasion could be assimilated to a colonisation process of defaunate areas as an inheritance of the last glaciation, with Eastern Europe being a provider of non-indigenous species and Western Europe as a recipient area. This process could be illustrated with the establishment of many Ponto-Caspian amphipod species in France. Based on the data provided by the Limnofauna Europea (Illies, 1978), 23 autochthonous species could be found in France, 7 of them inhabiting essentially brackish waters and 10 being restricted to small geographic areas in southern France. Thus, amphipod assemblages are essentially constituted of 6 species *versus* 63 in the Ponto-Caspian basin: this dissymmetry could be explained by an original pattern of amphipod invasions along a longitudinal European gradient from East to West. Another consequence of this geological

history of Western Europe is that the only solid basis to study changes in freshwater fauna is the oldest faunal list available for a given biota.

However, an acceleration is observed in non-indigenous species establishment within French aquatic ecosystems during the last few decades, suggesting that besides a natural process of colonisation, a human-mediated transport of species was essential and occurred in Europe (van der Velde *et al.*, 2000, bij de Vaate *et al.*, 2002). Nevertheless, the acceleration of this process could also be an artefact, related to the growing interest attached to biological invasions, and variations in sampling effort among years (Delisle *et al.*, 2003).

The identification and distribution of invasive species in France also lead to the difficult problem of species range extension assessment. Among them, some are naturally present within French hydrosystems, such as the Mediterranean *Proasellus coxalis* or *Atyaephyra desmaresti*. The difficulty in such cases consists in determining where the natural area of a species stopped, and where it should be considered as non-indigenous. It represents a very hard task because the spread can be very slow, and observations are generally only reported when species colonise rivers largely outside their natural range (i.e. once they have colonised another basin).

The 43 non-indigenous species identified represent more than one percent of the total macroinvertebrate French freshwater fauna. However, they could represent much more in terms of density (23% in species number, 88% in abundance in the Moselle River, a tributary of the Rhine River; Devin *et al.*, in press). In French aquatic ecosystems, non-indigenous crustaceans and molluscs have met with more success than other taxa. Such clumping of invaders occurs elsewhere (Morton, 1997; Van der Velde *et al.*, 2000) because closely related species tend to share traits that promote a successful introduction (Devin & Beisel, unpublished data). Moreover, molluscs and crustaceans are holobiotic species present all year round in the water and are generally represented in term of density and widely distributed in their natural ranges. These characteristics magnify their probability to be transported in a new ecosystem thanks to human activities (the human commensalism evoked by Ricciardi & Rasmussen, 1998).

This raises the problem of introduction vectors, not considered in the present analysis because precise data on vectors are scarce. For intercontinental transfers, aquatic species could mostly be transported by ballast water or for aquaculture purpose. For intracontinental transfers, besides these two vectors, recreational shipping and natural migration along newly opened waterways could also occur. However, while a mix between the different vectors should more probably be evoked, shipping and natural migration should constitute the most important vectors, in regard to the migratory corridor identified (bij de Vaate *et al.*, 2002).

Among non-indigenous species, six have been particularly successful and are of particular concern for freshwater biodiversity and ecosystem functioning: *Craspedacusta sowerbyi*, *Dreissena polymorpha*, *Corbicula fluminea*, *Dikerogammarus villosus*, *Chelicorophium curvispinum* and *Orconectes limosus*. They represent 13.9% of the non-indigenous fauna, a value that fits with the tens rule of Williamson (1996), stating that an average of one in ten successful introduced species will become invasive. The main characteristics of these species have already been detailed in several articles (van der Velde *et al.*, 2000, bij de Vaate *et al.*, 2002), as well as their ecological and economical impacts (van der Velde *et al.*, 2000, Westman, 2002, Simon & Townsend, 2003).

The main consequences of non-indigenous species in French freshwaters could be divided into several categories, and can lead to:

- a significant contribution in the oxygen budget of the river by filter activities of species such as the non-indigenous molluscs *D. polymorpha* and *Corbicula* sp. (Bachmann *et al.*, 1998)

- modifications of nutrient flux due to the filtration activity and pseudofeces production of bivalves, that cause the phenomenon of biodeposition (Botts *et al.*, 1996; MacIsaac, 1996; Hakenkamp *et al.*, 2001)

- transport and dissemination of parasites by species such as *Orconectes limosus* spreading the crayfish plague on colonised hydrosystems (Laurent, 1997; Van der Velde *et al.*, 2000)

- competition for space between the autochthonous amphipod *Gammarus pulex* and the two invasives *G. tigrinus* and *Dikerogammarus villosus* (Devin *et al.*, 2003)

- predation upon both native and non-indigenous fauna by introduced species with a predatory behavior such as *D. villosus* (Dick & Platvoët, 2000; Dick *et al.*, 2002)

- modification of the structural features of the habitats by fouling species, such as *D. polymorpha* and *Chelicorophium curvispinum* (Bachmann *et al.*, 2001) or ecosystem engineers such as crayfish (Statzner *et al.*, 2000).

This broad spectrum of impacts raises the need for predictive tools of ecosystem modifications in response to the establishment of non-indigenous species.

Finally, from a biological and ecological point of view, our results are in accordance with those of bij de Vaate *et al.* (2002): non-indigenous species are dominated by predators and filter feeders, and are generally thermophilous and euryhalines. The proportion of non-indigenous species among the different functional groups fits with their proportions among the main taxonomical units presented in the Table 2. However, species belonging to the groups ζ and γ_2 could be found among almost all taxonomical units. Thus, it seems that a high majority of non-indigenous species exhibit similar functional characteristics and that, more than a taxonomical clumping, a functional clumping of non-indigenous species take place, which evidenced that particular bio/ecological profiles could enhance invasiveness. Moreover, as previously discussed, most of the non-indigenous species come from the Ponto-Caspian basin. The complex and long evolution of this basin, with a highly variable salinity and temperature, and a high degree of endemism resulting from a long isolation, has led to a selection of euryecious species (Dumont, 1998; Reid & Orlova, 2002) presenting adequate features for a human-aided inoculation in new ecosystems.

A consequence of these similar functional characteristics of non-indigenous species is that, besides a threat to biodiversity in term of species composition, it could bring about functional redundancy (as defined by Rosenfeld, 2002) within macroinvertebrate communities.

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Table 1. List of 43 non-indigenous species that could be found among French hydrosystems. The functional groups given are from Usseglio-Polatera *et al.*, 2001.

	Espèce	Donor Area	Date of first observation	Functional group	References
Bryozoa	<i>Pectinatella magnifica</i> (Leidy, 1851)	North America	1994	ζ	1,2
Cnidaria	<i>Cordylophora caspia</i> Pallas, 1771	Ponto-Caspian	around 1970	ζ	3
	<i>Craspedacusta sowerbyi</i> Lankester, 1880	South-eastern Asia	1891	γ2	3,4,5,6
Turbellaria	<i>Dendrocoelum romanodanubiale</i> Codreanu, 1949	Ponto-Caspian	around 2002	δ2	7
	<i>Dugesia tigrina</i> (Girard, 1850)	North America	1950	γ2	3,7,8
Oligochaeta	<i>Branchiura sowerbyi</i> Beddard, 1892	South-eastern Asia	before 1926	ζ	3,9
Polychaeta	<i>Hypania invalida</i> (Grube, 1860)	Ponto-Caspian	1998	ζ	7, 10
Hirudinae	<i>Barbronia weberi</i> Blanchard, 1897	South-eastern Asia	around 2002		7
	<i>Caspiobdella fadejewi</i> Selensky, 1915	Ponto-Caspian	around 2002		7
Gastropoda	<i>Emmerica patula</i> Brumati, 1838	Central and Eastern Europe	1981		11
	<i>Lithoglyphus naticoides</i> (C. Pfeiffer, 1828)	Ponto-Caspian	1909	γ1	3, 12
	<i>Menetus dilatatus</i> (Gould, 1841)	North America	around 1980		11
	<i>Physella acuta</i> (Draparnaud, 1805)	North America	1862	γ1	3, 13
	<i>Potamopyrgus antipodarum</i> Gray, 1843	New-Zealand	around 1950	ζ	14
Bivalvia	<i>Anodonta cellensis</i> Schröter, 1779	Central and Eastern Europe	1985	ζ	15
	<i>Anodonta woodiana</i> (Lea, 1834)	South-eastern Asia	1985	ζ	15
	<i>Corbicula fluminalis</i> (Müller, 1774)	South-eastern Asia	1994	ζ	16
	<i>Corbicula fluminea</i> (Müller, 1774)	South-eastern Asia	1980	ζ	17,18
	<i>Dreissena polymorpha</i> (Pallas, 1771)	Ponto-Caspian	1847	ζ	19, 20
	<i>Musculium transversum</i> (Say, 1829)	North America	2000	ζ	21
Amphipoda	<i>Chelicorophium curvispinum</i> (Sars, 1895)	Ponto-Caspian	1993	ζ	3, 16, 22
	<i>Crangonyx pseudogracilis</i> Bousfield, 1958	North America	1995	ζ	3, 23, 22, 24
	<i>Dikerogammarus haemobaphes</i> (Eichwald,1841)	Ponto-Caspian	1998	ζ	7, 25
	<i>Dikerogammarus villosus</i> (Sovinski)Martinov, 1894	Ponto-Caspian	1997	ζ	7, 10, 24,23,22
	<i>Echinogammarus berilloni</i> (Catta, 1878)	Southern Europe	1937	ζ	3, 26
	<i>Echinogammarus ischnus</i> (Stebbing, 1906)	Ponto-Caspian	1997	ζ	7
	<i>Gammarus roeselii</i> Gervais, 1835	Central and Eastern Europe	1835	β	27
	<i>Gammarus tigrinus</i> Sexton, 1939	North America	1970'	β	3
	<i>Orchestia cavimana</i> Heller, 1865	Ponto-Caspian	1900-1930		3, 28
Isopoda	<i>Asellus aquaticus</i> (Linnaeus, 1758)	South-eastern Asia	Unknown	γ2	29
	<i>Jaera istri</i> Veuille, 1979	Ponto-Caspian	1999	ζ	7, 10
	<i>Proasellus coxalis</i> (Dollfus, 1892)	Southern Europe	1930	γ2	7, 29
Decapoda	<i>Astacus leptodactylus</i> Eschscholtz, 1823	Central and Eastern Europe	1890	γ2	30, 31
	<i>Atyaephyra desmaresti</i> (Millet, 1831-1832)	Southern Europe	1832	γ2	3, 30
	<i>Eriocheir sinensis</i> Milne-Edwards, 1853	South-eastern Asia	1930	γ2	32
	<i>Hemimysis anomala</i> Sars, 1907	Ponto-Caspian	2000		7, 27
	<i>Limnomysis benedeni</i> Czerniavsky, 1882	Ponto-Caspian	2000		7, 33
	<i>Orconectes limosus</i> (Rafinesque, 1817)	North America	1912	γ2	30, 32, 31
	<i>Pacifastacus leniusculus</i> (Dana, 1852)	North America	1973	δ2	30, 31, 34
	<i>Potamon ibericum tauricum</i> (Czerniavsky, 1884)	Southern Europe	1991	γ2	30, 35
	<i>Procambarus clarkii</i> (Girard, 1852)	North America	1976	δ1	30, 31
Diptera	<i>Aedes albopictus</i> Skuse, 1894	South-eastern Asia	1999	δ1	36
Trichoptera	<i>Pseudoneureclipsis lusitanicus</i> Malicky, 1980	Southern Europe	2001	ζ	37

1: d'Hondt & Condé, 1996; 2: Rodriguez & Vergon, 2002; 3: Dhur & Massard, 1995; 4: Parent, 1981; 5: Arvy, 1972; 6: Goy, 1971; 7: Anonymous, 2002; 8: Hoffman 1964; 9: Remy, 1926; 10: Devin *et al.* , in press; 11: Mouthon, 1986; 12: Remy, 1924; 13: Godron, 1862; 14: Boettger, 1954; 15: Girardill & Ledoux, 1989; 16: Bachmann *et al.*, 1997; 17: Fontan & Meny, 1997; 18: Mouthon, 1981; 19: Kinzelbach, 1992; 20: Léglize & Ollivier, 1981 ;21: Mouthon & Loiseau, 2000; 22: Lécureuil & Chovet, 2003; 23: Bollache *et al.*, 2004; 24: Vandenbossche, 2002; 25: Devin *et al.*, 2001; 26: Kinzelbach, 1972; 27: Jazdzewski, 1980; 28: Tetry, 1939; 29: Henry & Magniez, 1983; 30: Vigneux, 1997; 31: Laurent, 1997; 32: André, 1934; 33: Wittman & Ariani, 2000; 34: Vigneux *et al.* , 1993; 35: Charmantier, 1992; 36: Schaffner & Karch, 2000; 37: Tachet *et al.* , 2001

Table 2. Proportion of non-indigenous fauna among French macroinvertebrate fauna in freshwater systems. The total number of macroinvertebrate species in French freshwaters are unpublished results. Nematomorpha and parasites were excluded.

	Total number of species	Number of exotic species	Percentage of exotic species
Bryozoa	11	1	9.1
Porifera	5	0	0
Cnidaria	8	2	25
Platyhelminthes	33	2	6.1
Hirudinea	16	2	12.5
Oligochaeta	147	1	0.7
Polychaeta	1	1	100
Bivalvia	33	6	18.2
Gastropoda	55	5	9.1
Crustacea	86	21	24.4
Insecta	3257	2	0.06

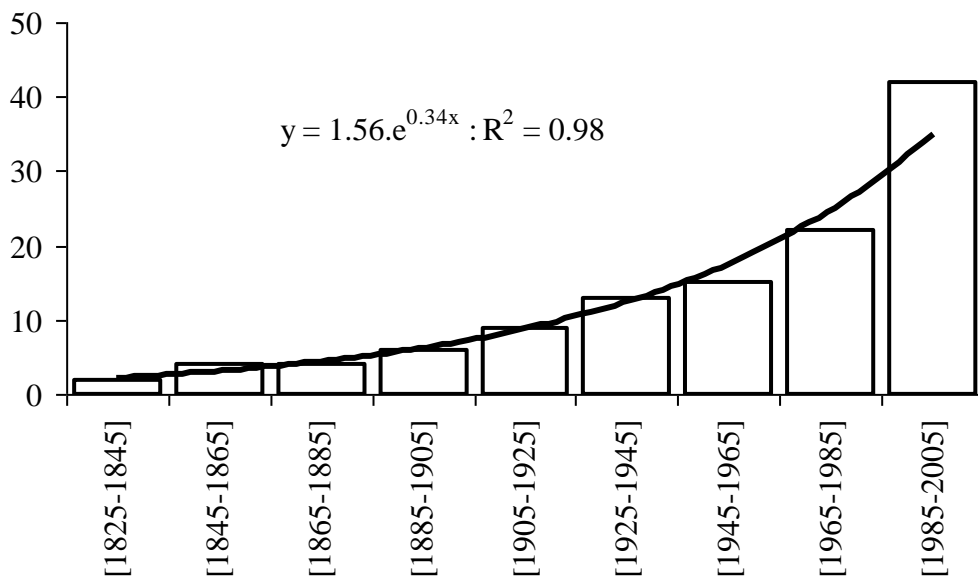


Figure 1: Trend in cumulative number of species established in French aquatic ecosystems. Three functions were tested, and the exponential one gave the best fit with our date. Only 42 species were considered in this analysis, as the date of arrival of *Asellus aquaticus* is not precisely known. Linear and power function relationships were also tested, but exhibit a lower correlation with the observed trend ($R^2=0.77$ and $R^2=0.89$ respectively).