

# A new understanding of somatic coliphages belonging to the Microviridae family in urban wastewater

Marion C Bichet, Marion Gardette, Benjamin das Neves, Julie Challant, Anaïs Erbs, Véronica Roman, Maëlle Robin, Stéphanie La Carbona, Christophe Gantzer, Nicolas Boudaud, et al.

## ▶ To cite this version:

Marion C Bichet, Marion Gardette, Benjamin das Neves, Julie Challant, Anaïs Erbs, et al.. A new understanding of somatic coliphages belonging to the Microviridae family in urban wastewater. Water Research, 2024, 249, pp.120916. 10.1016/j.watres.2023.120916. hal-04376223

## HAL Id: hal-04376223 https://hal.univ-lorraine.fr/hal-04376223v1

Submitted on 6 Jan 2024

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

A new understanding of somatic coliphages belonging to the *Microviridae* family in urban
 wastewater

4

#### 5 Authors:

- 6 Marion C. Bichet<sup>1,2</sup>, Marion Gardette<sup>1,2</sup>, Benjamin Das Neves<sup>2</sup>, Julie Challant<sup>2</sup>, Anaïs Erbs<sup>1,2</sup>,
- 7 Véronica Roman<sup>2</sup>, Maëlle Robin<sup>1</sup>, Stéphanie La Carbona<sup>1</sup>, Christophe Gantzer<sup>2</sup>, Nicolas
- 8 Boudaud<sup>1</sup> and Isabelle Bertrand<sup>2\*</sup>
- 9

### 10 Affiliations:

- <sup>11</sup> ACTALIA, Food Safety Department, F-50000 Saint-Lô, France.
- <sup>2</sup> University of Lorraine, CNRS, LCPME, F-54000 Nancy, France.
- 13

### 14 \*Corresponding author: Isabelle Bertrand

- LCPME, Campus Brabois Santé, Bâtiment B, 9 Avenue de la Forêt de Haye, BP 20199, F-
- 16 54505, Vandœuvre-lès-Nancy, France.
- 17 E-mail address: isabelle.bertrand@univ-lorraine.fr
- 18
- 19

20 ABSTRACT

Somatic coliphages (SC) and F-specific RNA coliphages (FRNAPH) have been included in 21 regulations or guidelines by several developed countries as a way of monitoring water safety 22 and the microbiological quality of shellfish harvesting waters. SC are highly diverse in their 23 morphology, size and genome. The *Microviridae* family contains three genera of phages 24 25 (Alphatrevirus, Gequatrovirus, and Sinsheimervirus), all having a capsid of similar morphology (icosahedral) and size (25-30 nm in diameter) to that of common pathogenic enteric viruses. 26 27 Three PCR assays specific for each genus of *Microviridae* were designed to study these phages in raw and treated wastewater (WW) in order to gain knowledge about the diversity and 28 prevalence of Microviridae among SC, as well as their inactivation and removal during WW 29 treatments. Among the four wastewater treatment plants (WWTPs) monitored here, two 30 31 WWTPs applied disinfection by UV light as tertiary treatment. First, we noticed that Microviridae represented 10 to 30 % of infectious SC in both raw and treated WW. 32 Microviridae appeared to behave in the same way as all SC during these WW treatments. As 33 expected, the highest inactivation, at least 4 log<sub>10</sub>, was achieved for infectious Microviridae 34 and SC in both WWTPs using UV disinfection. PCR assays showed that the highest removal of 35 *Microviridae* reached about 4 log<sub>10</sub>, but the phage removal can vary greatly between WWTPs 36 using similar treatments. This work forms the basis for a broader evaluation of Microviridae 37 38 as a viral indicator of water treatment efficiency and WW reuse.

39

40

41 Keywords: Somatic coliphages; *Microviridae*; Wastewater; Viral pollution; Bacteriophages
 42

43 **1. Introduction** 

Waterborne enteric viruses are a major threat to global public health. These viruses 44 are transmitted by the faecal-oral route and most of them are persistent in the environment 45 (Kotwal and Cannon, 2014). Metagenomic studies have demonstrated the abundance and 46 diversity of enteric viruses in the human gastrointestinal tract (Reyes et al., 2012), as well as 47 48 in urban wastewater (WW) (Aw et al., 2014; Cantalupo et al., 2011; Petrovich et al., 2019). Enteric viruses may be discharged in aquatic environments depending on both the treatments 49 applied in WWTP and their initial concentration. Viral contamination of environmental waters 50 has been involved in drinking waterborne outbreaks (Moreira and Bondelind, 2017), 51 foodborne outbreaks due shellfish consumption such as oysters (Bellou et al., 2013), and 52 direct human contaminations during recreational activities (Sinclair et al., 2009). It is therefore 53 54 important to monitor viral pollution in environmental waters for more effective viral hazard management. However, the direct detection of pathogenic enteric viruses in environmental 55 waters remains difficult owing to irregular sampling, the small volume of water analysed, the 56 temporal variability, and the low viral content outside of epidemic peaks. For this reason, F-57 specific RNA coliphages (FRNAPH) and somatic coliphages (SC) have been included in 58 regulations or guidelines by numerous developed countries as a way of monitoring water 59 quality (US EPA, 2006; US EPA, 2015; European Commission, 2020; Health Canada, 2019) and 60 61 ensuring the safety of shellfish (ASQAP, 2022; US FDA, 2019). The main advantages of FRNAPH and SC are their high prevalence and concentration in urban WW – with between 5 and 8 log<sub>10</sub> 62 plaque forming units (PFU)/L (Blanch et al., 2006; Cole et al., 2003; Farkas et al., 2020; 63 Gourmelon et al., 2007; Hartard et al., 2015; Havelaar et al., 1990; Korajkic et al., 2020) and 64 the standardised methods used, which allow a fast and accurate quantification of the number 65 of infectious particles (ISO 10705-1, 1995; ISO 10705-2, 2001). Furthermore, the effectiveness 66

of the removal of SC and FRNAPH during WW treatments has been reported to be similar to
that of pathogenic enteric viruses(Carducci et al., 2009; Martín-Díaz et al., 2016). SC have been
shown to be more abundant than FRNAPH, excepted in some situations such as reclaimed
water, in which UV is the main factor in viral inactivation (Jofre et al., 2016).

SC are non-enveloped phages whose taxonomy has recently been modified. The three 71 72 families of tailed phages Myoviridae, Podoviridae, and Siphoviridae have been removed from the phage taxonomy proposed by the International Committee on Taxonomy of Viruses (ICTV) 73 because they are not monophyletic (Turner et al., 2021). Now, these phages belong to the 74 class of Caudoviricetes and the terms myovirus, podovirus, and siphovirus have been 75 suggested for publications to indicate their morphology (Turner et al., 2021). Due to their high 76 genomic diversity, there are not currently any available methods for the quantification of SC 77 78 genomes. It is therefore interesting to focus on one specific family; the SC belonging to the Microviridae family are morphologically closer to the main enteric pathogenic viruses in terms 79 of capsid shape (icosahedral) and size (around 25 nm) than the other SC. They are organised 80 into three genera: Alphatrevirus, Gequatrovirus and Sinsheimervirus; the coliphages  $\alpha$ 3, G4, 81 82 and  $\phi$ X174 were chosen as representative of these genera. A previous study, using transmission electron microscopy on infectious phages, observed that Microviridae were less 83 removed by a full-scale membrane bioreactor (MBR) than the other SC (Purnell et al., 2015). 84 85 The authors suggested that the highest susceptibility of the tailed SC to MBR could be due to adsorption to solids and/or damage within the MBR process. Since only *Microviridae* were 86 observed among infectious SC in treated WW, the viral morphology of SC may be an important 87 factor in their removal by MBR membranes (Purnell et al., 2015). It has also been shown that 88 the removal of  $\phi$ X174 genomes after WW treatment is similar to that of some other 89

pathogenic viruses commonly found in WW like the adenovirus or polyomavirus (Sidhu et al.,
2017).

First, this work aimed to study the quantity and diversity of Microviridae compared to 92 SC as a whole in raw and treated wastewater. The second objective was to evaluate the 93 inactivation and removal of these Microviridae during different WW treatments. To do this, 94 95 new PCR assays were developed for each genus of *Microviridae*. For the purposes of this study, the three genera will be named  $\alpha$ 3-like, G4-like and  $\varphi$ X174-like phages, for Alphatrevirus, 96 Gequatrovirus and Sinsheimervirus, respectively. The PCR assays were carried out on raw and 97 treated WW samples to quantify the number of genomes present as well as to genotype the 98 infectious PFUs particles to estimate the abundance of *Microviridae* as a proportion of all SC. 99

100

#### 101 **2. Materials and methods**

102 2.1. Water samples

Influents and effluents of urban WW from four different French WWTPs were selected 103 to consider the variability of urban WW. In this study, we will refer to them as WWTP-1, 104 WWTP-2, WWTP-3, and WWTP-4 (Figure 1A). Influent samples were collected at the WWTPs' 105 entry point and effluent samples were collected just before environmental discharge. All 106 samples were collected during 24h periods using automated samplers. The characteristics of 107 108 the WWTPs (location, treatment capacity, type of WW, automated sampler, steps implemented to treat WW) are described in Table 1. Between 8 and 12 independent samples 109 of influent and effluent WW were collected in each WWTP with a total of 38 influent samples 110 and 34 effluent samples (see Supplementary Table S1). For WWTP-1 and WWTP-2, samples 111 were collected between September 2021 and April 2022 and between April 2022 and May 112

2022, respectively. For WWTP-3 and WWTP-4, samples were collected between September
2021 and June 2022.

115

116 2.2. Production, quantification, and genotyping of infectious somatic coliphages

Three representative *Microviridae* strains, φX174, α3, and G4-like, belonging 117 118 respectively to the Sinsheimervirus, Alphatrevirus and Gequatrovirus genera, were selected for the development of specific real-time PCR methods. Referenced strains were used for the 119  $\varphi$ X174 phage (ATCC 13706) and  $\alpha$ 3 phage (HER 536, Félix d'Hérelle Reference Center for 120 Bacterial Viruses from the University of Laval, Québec, Canada), whereas the G4-like phage 121 was isolated from wastewater and fully sequenced (as described below). Each phage was 122 separately replicated and produced according to the ISO 10705-2 standard method (ISO 123 124 10705-2, 2001) with minor adjustments (the chloroform step was not carried out and the bacterial host used was *Escherichia coli* CN13, which is also known as WG5 (ATCC 700078). 125 Phage suspensions were then centrifuged (6000 × g for 20 min at 4 °C) and supernatants were 126 filtered through 0.22-µm sterile syringe filters made of hydrophilic polyethersulfone (Cytiva, 127 USA). Initial concentrations of viral stocks were determined using the ISO 10705-2 standard 128 method (NF EN ISO 10705-2, 2001). Only infectious SC can be replicated using the E. coli CN13 129 bacterial strain, since the absence of F-pilus on the surface of the bacterial strain prevents the 130 131 replication of infectious FRNAPH. Concentrations of phage suspensions were about 10<sup>12</sup> PFU/mL for the  $\alpha$ 3 and G4-like phages, and about 10<sup>8</sup> PFU/mL for  $\phi$ X174 phages. The phage 132 suspensions were stored at 4 °C for up to two months. 133

134 Infectious SC were enumerated in duplicate in each WW sample using the ISO 10705-135 2 standard method as described above. After enumeration, plates containing 10 to 70 PFUs 136 per sample were selected for genotyping. All the PFU of each plate were then individually

recovered using sterile loops and resuspended in 1 mL of PBS-5% Glycerol by vortexing for 20
sec. The total PFU analysed for the four WWTP reached 1,001 for influents and 664 for
effluents (Table S1). Samples were then stored at -20 °C before DNA extraction for up to two
weeks.

141

#### 142 2.3. Viral DNA extraction

Viral DNA extraction was directly performed on one mL of WW. Two mL of NucliSENS® 143 lysis buffer (BioMérieux, France) were added to the sample, followed by vortexing and 144 incubation at room temperature for 10 min. Samples were then stored at -20 °C for up to 145 seven days or immediately extracted using NucliSENS® magnetic extraction reagents on 146 NucliSENS<sup>®</sup> MiniMAG (BioMérieux, France) according manufacturer's to the 147 148 recommendations. A negative control (water) was included in each run of extraction. The extracted DNA was eluted in 100 µL of elution buffer, purified with a OneStep PCR Inhibitor 149 Removal Kit (Zymo Research, Ozyme, France) according to the manufacturer's 150 recommendations. Briefly, a column was prepared prior to use with a solution provided in the 151 152 kit. The matrix of the column has been designed for efficient removal of polyphenolic compounds, humic and fulvic acids, tannins, etc. The sample was transferred on the matrix 153 and recovered by centrifugation. The samples were stored at -80 °C for up to one month 154 155 before PCR analyses.

The individual plaques recovered in PBS-5% Glycerol were heated at 95°C for 5 min at 500 rpm to extract DNA (ThermoMixer comfort, Eppendorf, France). Then, the genotype of each plaque was determined by qPCR as described below. All plaques containing a mix of genotypes were removed from the analysis. This represented 2% of the total number of plaques collected over all samples from the four WWTPs (see Supplementary Table S1).

161

#### 162 2.4. Design of primers and TaqMan probes

The primers and probes were designed using Primer3Plus and the genomic sequences 163 available on Genbank (in 2021, when the primers were designed) for the three genera of the 164 *Microviridae* family. Using BLAST alignment for nucleotides, the identity ranged from 98.9% 165 166 to 99.9% (coverage: 100%) between  $\phi$ X174 sequence with accession number NC 001422.1 and the 114 other complete sequences of *Sinsheimervirus*. The identity ranged from 84.0% to 167 100% (coverage: 84% to 100%) between  $\alpha$ 3 sequence with accession number NC 001330.1 168 and the 14 other Alphatrevirus sequences. It ranged from 91.3% to 100% (coverage: 99% to 169 100%) between G4 sequence with accession number NC\_001420.2 and the 19 other 170 Gequatrovirus sequences. The previously described PCR system for  $\varphi$ X174 detection (Dika et 171 172 al., 2015) matched with only 63% of the 115 sequences of Sinsheimervirus, but this result could be improved to 86% by degenerating one nucleotide on the reverse primer at position 1727 173 on NC\_001422.1. The system with the degenerated reverse primer was used for  $\phi$ X174-like 174 phage detection and quantification. The primers and probes were located in the replicase 175 gene for  $\alpha$ 3-like phages and in the hypothetical Protein A for G4-like phages because of the 176 relatively small amount of sequence variation encountered in those areas. BLAST alignments 177 showed that these two PCR systems detected all the complete sequences of their genus 178 179 disponible on Genbank. The specificity for each genus could also be estimated at 100% for the three PCR assays. An additional validation was done by testing DNA from  $\alpha$ 3, G4-like, and 180 φX174 with each PCR detection assay. The specifications of the oligonucleotides are given in 181 Supplementary Table S2. 182

183

184 2.5. Digital droplet PCR

185 The digital droplet PCR (ddPCR) technology was used to quantify the genomes of each Microviridae genus directly in the WW samples. It was carried out using a ddPCR supermix for 186 Probes (no dUTP) kit with a QX200 ddPCR system (Bio-Rad). Each reaction contained a final 187 concentration of 1 X Supermix, 900 nM of each primer, 130 nM of the  $\phi$ 174-like probe, or 230 188 nM of the  $\alpha$ 3-like and G4-like probes (Eurogentec, Belgium) and molecular-grade water. 189 190 Template DNA (5.5  $\mu$ L) was added for a final reaction volume of 22  $\mu$ L. Droplet generation was performed according to the manufacturer's instructions (BioRad). The resulting droplets were 191 192 then transferred to a Thermal Cycler for PCR amplification using the following parameters with a ramp rate of 70% between each step: 95 °C for 10 min, 40 cycles of 94 °C for 30 sec and 55 193 °C for 1 min, followed by 98 °C for 10 min and a final step at 4 °C for 30 min. The droplets were 194 kept at 4 °C up to 12 hours before analysis using the QX200 Droplet Reader (BioRad). 195

The limit of quantification (LOQ) of the ddPCR assay was of three positive droplets for 20,000 total droplets. By modelling with a Poisson distribution, this represented a concentration ~50 genome copies (GC)/100  $\mu$ L of extracted DNA. As only one mL of WW was analysed, the LOQ was estimated to be ~4.7 log<sub>10</sub> GC/L for these experiments.

200

#### 201 2.6. Real-time PCR

The real-time PCR technique was used to genotype the PFU recovered from the plates for each *Microviridae* genus. For influents samples, a total of 1,001 PFU were genotyped. The number of genotyped PFU per WWTP varied between 218 (WWTP-1) and 283 (WWTP-3). For effluent samples, a total of 664 PFU were genotyped. The number of genotyped PFU per WWTP varied between 110 (WWTP-4) and 269 (WWTP-2). The detailed data are given in Supplementary Table S1. The reactions were performed using the TaqMan Universal PCR Master Mix (ThermoFisher Scientific). In accordance with the manufacturer's

209 recommendations, each reaction was performed in a 25 µL reaction mixture containing 5 µL of template DNA and 15 µL of 1 × TaqMan mix, 800 nM of each primer and 200 nM of the 210 probe, and molecular grade water (Sigma). The amplification was carried out on a CFX96 211 Touch Real-Time PCR System (BioRad) using the following parameters: 50 °C for 2 min, 95 °C 212 for 10, min, followed by 45 cycles of 95 °C for 15 sec and 60 °C for 1 min. Infectious phage 213 214 genotyping was performed using the extracted phage plaque DNA and was considered positive when Cq was below 30 because a PFU is expected to contain a high quantity of 215 genome. This limit corresponded to  $\sim 5 \log_{10} \text{GC/PFU}$ . 216

217

218 2.7. Isolation and sequencing of environmental G4-like phage

A phage belonging to G4-like phages was isolated from an influent sample collected at 219 220 WWTP-1 during SC enumeration and genotyping of PFU, amplified on E. coli CN13 and quantified at  $2 \times 10^{13}$  GC/mL before sequencing analysis. Several steps were then applied to 221 obtain a sufficient concentration of purified phage DNA for sequencing analysis. First, this 222 phage was amplified on a plate; then the top agar layer was separated from the other layer 223 224 starting from the edge of the plate using sterile forceps and transferred to 200 mL of PBS, 225 homogenised by vortexing and then slowly agitated during 4 hours at room temperature. After centrifugation at 2,520  $\times$  g for 10 min at 4 °C, the supernatant was filtered through 5-µm and 226 227 0.22- $\mu$ m filters. The filtrate was ultra-centrifuged at room temperature, at 53,900  $\times$  g, for 7 hours (Beckman Optima XE-90; Rotor type 19), and the resulting pellet was resuspended in 4 228 mL of PBS. Before sequencing, the nucleic acids were extracted using NucliSENS® magnetic 229 extraction reagents on NucliSENS® MiniMAG (BioMérieux, France) and treated with DNase-I 230 to remove bacterial DNA. The bacterial DNA removal was validated using a 16S PCR method 231 (Muyzer et al., 1993; Nadkarni et al., 2002). As SC are DNA phages, the treatment with DNAse 232

233 resulted in a 3 log<sub>10</sub> decrease in phage DNA, but  $2 \times 10^{13}$  GC/mL were still present in the sample. The phage was sequenced using next generation sequencing (NGS) (Microsynth AG, 234 Balgach, Switzerland). Briefly, the second-strand DNA was synthesised with DNA polymerase 235 I, an Illumina TruSeq DNA library was prepared and the sequencing was performed using 236 Illumina MiSeq 1 Mio reads (2x250 bp runs, v2). After demultiplexing and trimming of Illumina 237 238 adaptor residuals, the assembled sequence was obtained using Unicycler v0.5.0. The sequence is available on GenBank with accession number OR492646. BLAST analysis showed between 239 92.1% and 96.7% identity (coverage: 100%) with the 20 other complete sequences of G4-like 240 phages, 77% identity (coverage: 33%) with the NC\_001422.1 sequences (Escherichia phage 241 phiX174, complete genome) and between 71.1% and 75.8% identity (coverage: 24% to 32%) 242 with the 15 complete sequences of  $\alpha$ 3-like phages. 243

244

245 2.8. Statistical analysis

All the statistical tests were performed using GraphPad-PRISM software (version 246 9.5.0). The  $\alpha$  level was fixed at 0.05. All data were analysed for single outliers using the Grubbs 247 test. After removing the detected single outliers, the normality of the data sets was assessed 248 using the Shapiro-Wilk test. All data were independent of each other. So, we used the Kruskal-249 Wallis test followed by a post-hoc Dunn multiple pairwise comparison test to compare more 250 251 than two samples. A Mann-Whitney test was used to compare two samples. Asterisks were used to indicate significant differences: \* for p-value  $\leq 0.05$ ; \*\* for p-value  $\leq 0.01$ ; \*\*\* for p-252 value  $\leq$  0.001, and \*\*\*\* for p-value  $\leq$  0.0001. 253

254

#### 255 **3. Results**

256 3.1. Assessment of the presence and diversity of *Microviridae* in raw wastewater

257 Our first aim was to assess the presence and diversity of the three genera of 258 *Microviridae* in raw WW by detecting their corresponding infectious particles and viral 259 genomes. Thirty-eight influent samples from four WWTP locations in two different regions of 260 France were analysed (Figure 1A and Supplementary Table S1).

The three specific genera of *Microviridae* were detected using three specific PCR assays 261 262 designed for  $\alpha$ 3-like, G4-like and  $\phi$ X174-like phages. The specificity of the PCR assays, evaluated using BLAST alignments (see above in 2.4), was confirmed by the absence of cross-263 reactivity when testing DNA from  $\alpha$ 3, G4-like, and  $\phi$ X174 with each PCR detection system. 264 Both qPCR and ddPCR methods could be used to quantify the Microviridae genomes in 265 environmental samples. For convenience, the qPCR method was used for determine which of 266 the three genera of *Microviridae* the plaques of infectious SC were made up of, while the 267 268 ddPCR method was used for the direct quantification of the genome copies in WW samples.

Infectious SC were detected using the standardised PFU method (NF EN ISO 10705-2, 269 2001) in all the WW samples. The mean concentrations of infectious SC in influents reached 270 7.1 ± 0.5 log<sub>10</sub> PFU/L (WWTP-1), 7.4 ± 0.2 log<sub>10</sub> PFU/L (WWTP-2), 8.1 ± 0.4 log<sub>10</sub> PFU/L (WWTP-271 3) and 8.0  $\pm$  0.6 log<sub>10</sub> PFU/L (WWTP-4) (Figure 1B), with the SC concentrations in WWTP-1 272 being significantly lower than those observed in WWTP-3 (p-value = 0.0009) and WWTP-4 273 (p-value = 0.0064). After genotyping each plaque from each sample, the results allowed us to 274 275 determine which of the three Microviridae genera (or other SC) each PFU was made up of (see Supplementary Figure S1B). The mean concentration of infectious Microviridae for each site 276 varied from 5.1  $\pm$  2.4 log<sub>10</sub> PFU/L (WWTP-1) to 7.4  $\pm$  0.5 log<sub>10</sub> PFU/L (WWTP-4) (Figure 1B). As 277 observed for infectious SC, the concentration of infectious Microviridae was significantly lower 278 at WWTP-1 than at WWTP-3 (p-value = 0.0031) and WWTP-4 (p-value  $\leq$  0.0001). From these 279 values, we estimated that infectious Microviridae make up on average between 12 % ± 9 % 280

281 (WWTP-1) and 30 % ± 14 % (WWTP-4) of total infectious SC in influents. If we take into account all four WWTPs, the mean concentrations of infectious *Microviridae* of each genus 282 varied from 2.6  $\pm$  3.4 log<sub>10</sub> PFU/L for G4-like phages to 4.0  $\pm$  3.2 log<sub>10</sub> PFU/L for  $\alpha$ 3-like phages 283 (data not shown). When analysing each site separately (Figure 1C), the only significant 284 difference was a lower proportion of G4-like phages than  $\alpha$ 3-like phages in influents from 285 286 WWTP-2 (p-value = 0.0168). Nevertheless, the wide distribution of the concentration values 287 with negative samples frequently observed for at least one genus may hide the differences between genera. 288

The mean concentration of Microviridae genomes varied from 7.6 ± 0.3 log10 GC/L 289 (WWTP-1) to 8.1 ± 0.3 log<sub>10</sub> GC/L (WWTP-3) (Figure 1B). As observed for the infectious phages, 290 WWTP-1 had a lower concentration of *Microviridae* genomes than WWTP-3 (p-value = 0.0149) 291 292 and WWTP-4 (p-value = 0.0288). If we take into account all four WWTPs, the mean concentrations of each genus varied from  $6.7 \pm 0.6 \log_{10}$  GC/L for G4-like phages to  $7.6 \pm 0.4$ 293  $\log_{10}$  GC/L for  $\alpha$ 3-like phages (data not shown). When analysing each site separately (Figure 294 1D), the concentrations of  $\alpha$ 3-like genomes were significantly higher than those of G4-like 295 296 genomes at all WWTPs (p-values from ≤0.0001 for WWTP-2 and -4 to 0.0316 for WWTP-1). The concentrations of the  $\phi$ X174-like genomes were in between those of the two other 297 genera, with no significant differences. These original genome data may be more 298 299 representative of the real proportions of  $\alpha$ 3-like, G4-like and  $\phi$ X174-like phages than those obtained for PFUs as it was observed less variability for any given site for the genome 300 parameter than what we observed for the infectious parameter. 301

302

303 3.2. Comparison of the behaviour of *Microviridae* and somatic coliphages during wastewater
 304 treatments

For the four same WWTPs, 34 effluent samples were collected in parallel with influent samples and analysed using the same methods to determine the concentrations of infectious SC and *Microviridae* and *Microviridae* genomes at the end of WW treatment (see Supplementary Table S1 and Supplementary Figure S1). From influent and effluent concentrations, the removal of infectious SC and *Microviridae* as well as the removal of *Microviridae* genomes could be estimated and expressed as a log<sub>10</sub> reduction (Figure 2A).

A reduction of  $1.9 \pm 0.3 \log_{10}$  and  $2.9 \pm 0.2 \log_{10}$  was observed for infectious SC at WWTP-1 and -2. WWTP-3 and -4 achieved a reduction of  $4.2 \pm 0.4 \log_{10}$  and  $3.9 \pm 0.5 \log_{10}$ , respectively, with both sites doing tertiary disinfection treatment based on UV light (Figure 2A and Table 1). The reduction in infectious SC was statistically higher in WWTP-3 and -4 than in WWTP-1 (p-value  $\leq 0.0001$  for WWTP-3 and p-value = 0.0002 for WWTP-4). No significant difference was observed between WWTP-2 and the three other WWTPs.

The reduction in infectious Microviridae followed a similar trend to that observed for 317 all SC (Figure 2A), but higher standard deviation values were noticed for Microviridae than for 318 SC. The reduction of infectious *Microviridae* reached  $0.5 \pm 2.5 \log_{10}$  at WWTP-1 and  $3.1 \pm 1.5$ 319 320  $log_{10}$  at WWTP-2, while average reductions of 4.0 ± 3.6  $log_{10}$  and 5.6 ± 1.9  $log_{10}$  were observed at WWTP-3 and WWTP-4, respectively. As observed for all SC, the reduction in infectious 321 Microviridae was significantly lower at WWTP-1 than at WWTP-3 (p-value = 0.0042) and 322 323 WWTP-4 (p-value  $\leq$  0.0001). Based on the concentration values of infectious particles (Supplemental Figure 1), the average proportions of infectious *Microviridae* were estimated 324 to be between 17 % ± 17 % (WWTP-3) and 32 % ± 16 % (WWTP-2) of all infectious SC in 325 effluent samples. Overall, except for WWTP-1 where infectious Microviridae were significantly 326 less reduced than infectious SC (p-value = 0.0088), similar behaviour was observed between 327

infectious SC and *Microviridae* during WW treatments, even when tertiary UV treatment was
 carried out.

The removal of *Microviridae* particles was estimated using genome quantification data 330 (Figure 2A). The highest mean removal of *Microviridae* particles was 4.2 ± 2.1 log<sub>10</sub> for WWTP-331 2 and was significantly higher than the reduction of about 2.0 log<sub>10</sub> obtained at the three other 332 333 WWTPs (p-values = 0.0029, 0.0092, and 0.0417, respectively). Despite both applying similar 334 steps for the entire wastewater treatment, WWTP-1 and -2 showed a significant difference (pvalue = 0.0029) in effectiveness in removing *Microviridae* particles (about  $2 \log_{10}$ ). Even if the 335 treatments applied in these two WWTPs were similar, they may differ in their conditions of 336 application. The 4.0 log<sub>10</sub> reduction obtained for infectious *Microviridae* and SC at WWTP-3 337 and -4 was not observed for the *Microviridae* genomes detected with a single target. 338

339 Few differences were found between the three genera of Microviridae. For infectious Microviridae (Figure 2B), we did not observe a reduction in the concentration of G4-like 340 phages at WWTP-1 and WWTP-2 because infectious G4-like phages were more frequently 341 detected in effluents than in influents in these two WWTPs. Thus, for these two WWTPs, G4-342 like phages represented only 11.5% and 8.9% of infectious *Microviridae* in influents, whereas 343 they represented 56.8% and 63.5% of infectious Microviridae in effluents (Table S1). This led 344 to the only significant difference observed between the three genera for infectious 345 346 *Microviridae* with a higher removal of  $\alpha$ 3-like phages than G4-like phages at WWTP-2 (p-value = 0.0251). If we take into account all the data, irrespective of the WW treatment 347 method used, there was similar reductions of infectious particles for all three genera (p-value 348 > 0.05). Regarding genome data (Figure 2C), the only difference between genera was observed 349 at WWTP-3, where the decrease in  $\alpha$ 3-like phages was higher than those observed for G4-like 350 phages (p-value = 0.0063) and  $\varphi$ X174-like phages (p-value = 0.0126). If we take into account 351

all the data, there was significantly higher reduction of  $\alpha$ 3-like phages than  $\varphi$ X174-like phages during WW treatments (p-value = 0.0186). Moreover, the higher genome concentrations of  $\alpha$ 3-like phages relative to G4-like phages in all influent samples were no longer visible in the effluent samples (see Supplementary Figure 1C). Despite the variability of the data, especially for infectious *Microviridae*, our results suggest that the removal of  $\alpha$ 3-like phages during WW treatments was slightly more efficient than for the two other genera (G4-like and  $\varphi$ X174-like).

#### 359 4. Discussion

This study aimed to increase knowledge about SC belonging to the *Microviridae* family 360 in WW, and to better understand the effect of WW treatments on their inactivation and 361 removal. PCR assays were developed here for the quantification of the three genera of 362 363 *Microviridae*, which is the only remaining recognized family in the SC (Turner et al., 2021). The design of these PCR assays was done by using the complete sequences available on Genbank 364 which are more numerous for  $\phi$ X-174-like genus than the other genera. Three PCR assays 365 were necessary since this family is too genetically diverse to create a single PCR assay that can 366 367 target all its members. A previous study proposed a PCR assay that only targeted the φX-174like genus (Lee, 2009) and did not detect all Microviridae or compare the three genera of 368 Microviridae between each other. By conducting our PCR assays on PFUs, we also quantified 369 370 infectious Microviridae in WW, which provided information on both the removal and inactivation of Microviridae by WW treatments. Raw and treated WW samples were collected 371 in four French WWTPs using tertiary treatments based either on lamellar settling for the 372 removal of phosphorous and residual suspended solids or on UV disinfection. 373

All influent samples were positive for infectious SC using the standardised method with values ranging from 7.1 to 8.2 log<sub>10</sub> PFUs per litre. A meta-analysis of infectious SC in raw WW

376 has previously shown a significant difference in the mean density distributions between geographic regions (Nappier et al., 2019). Nevertheless, our values were consistent with mean 377 concentration values ranging from 5.1 to 9.7 log<sub>10</sub> PFUs per litre previously reported for raw 378 WW (Blanch et al., 2006; Chacón et al., 2020; Jofre et al., 2021; Kelmer et al., 2023; Levantesi 379 et al., 2010; Lucena et al., 2004; Muniesa et al., 1999; Purnell et al., 2015; Truchado et al., 380 381 2021; Zhang and Farahbakhsh, 2007). Moreover, a mean concentration value of 7.1 log<sub>10</sub> PFUs per litre had been previously obtained for WWTP-1 featured in this study (Lucena et al., 2004). 382 The reduction in infectious SC between influents and effluents reached 2 and 3 log<sub>10</sub> in 383 WWTP-1 and -2, respectively. These WWTP included tertiary treatment for phosphorous 384 removal based on flocculation with FeCl<sub>3</sub> and lamellar setting. About 2.7 log<sub>10</sub> removal of 385 infectious SC was observed in a WWTP applying similar treatment including flocculation (n=14 386 387 samples) (Fu et al., 2010). This tertiary treatment was estimated to eliminate about 0.5 to 0.8 log<sub>10</sub> of infectious SC (Fu et al., 2010; Gantzer et al., 1998; Lucena et al., 2004). The reduction 388 in infectious SC between influents and effluents reached 4 log<sub>10</sub> in WWTP-3 and -4. It can be 389 considered that the difference of 1 to 2 log<sub>10</sub> observed between WWTP-1 and -2 on one hand 390 391 and WWTP-3 and -4 on the other hand was due to UV disinfection because it was the main difference between these WWTPs. Similarly, Korajkic and colleagues (2022) observed that UV 392 light treatment improved infectious SC reduction by 1 to 2 log<sub>10</sub>. About 5 log<sub>10</sub> reduction of 393 394 infectious SC was observed in two WWTPs using UV or chlorine as tertiary treatment (Truchado et al., 2021). 395

The three PCR assays developed here allowed us to measure the genome concentrations of each *Microviridae* genus and the concentrations of infectious *Microviridae*. The *Microviridae* genomes were systematically detected in raw WW and their concentrations were between 7.6 and 8.1 log<sub>10</sub> GC per litre. By using the PCR assay designed by Lee (2009),

400 the genome of the  $\phi$ X-174-like genus was detected at 6 to 7 log<sub>10</sub> GC per litre of raw WW in Australia (Sidhu et al., 2018, 2017). We observed a predominance of the  $\alpha$ 3-like genome 401 compared to the G4-like genome in raw WW, indicating differences in the initial 402 concentrations of the three genera of Microviridae. Infectious Microviridae were also 403 systematically detected in raw WW and they represented between 12 % and 30 % of all 404 405 infectious SC (21% by considering the 1,001 PFU from all influents samples together). 406 Moreover, we analysed all PFU from plates containing between 10 and 70 PFU in order to have more chance to detect phages present at low frequency. Our percentage values for 407 Microviridae were considerably higher than those observed by Muniesa et al., (1999), who 408 detected no more than 2 % of Microviridae among infectious SC in raw WW. However, these 409 authors analysed plates containing no more than 20 PFU. This low quantity of PFU per plate 410 411 may have contributed to the low detection of *Microviridae* as suggested by previous study performed on bacterial diversity (Bianchi and Bianchi, 1982). Metagenomic approaches 412 previously applied to WW gave rates that were closer to ours, with Microviridae representing 413 19 % to 37 % of total SC in raw sewage (Cantalupo et al., 2011; Aw et al., 2014). 414

The analysis of raw and treated WW with our PCR assays showed the level of removal 415 of Microviridae, whereas PCR assays on PFUs provided a relative information about the rate 416 of Microviridae inactivation. PCR assays allow the quantification of both infectious and non-417 418 infectious Microviridae and several states of viral particles (infectious, non-infectious with intact capsid, non-infectious with altered capsid) can be present and detected using PCR 419 methods in environmental samples (Randazzo et al., 2018; Wurtzer et al., 2021). For the two 420 WWTPs that used tertiary settling treatments, the reduction in Microviridae genomes (2 to 4 421 log<sub>10</sub>) was greater than the reduction in infectious *Microviridae* (0.5 to 3 log<sub>10</sub>), showing that 422 physical removal of Microviridae particles predominated over inactivation. This could be due 423

424 to the adhesion of a part of Microviridae phages to suspended matter leading to their removal with these particulates during flocculation and lamellar settling. The study of the surface 425 properties of Microviridae, especially their relative hydrophobicity as performed for other 426 phages (Dika et al., 2015), could contribute to the understanding of this phenomenon. On the 427 contrary, for tertiary treatment using UV disinfection, the reduction in infectious Microviridae 428 429 (4 to 5  $log_{10}$ ) was greater than the reduction in *Microviridae* genomes (2  $log_{10}$ ). This was expected, because viral inactivation is due to genomic lesions during UV disinfection (Beck et 430 al., 2016; Pecson et al., 2011; Simonet and Gantzer, 2006), but damage due to UV are often 431 limited to some parts of the genome and the use of a single genome target is commonly no 432 sufficient to evaluate such a treatment (Young et al., 2020). As we targeted only a small part 433 of Microviridae genome (~100 nucleotides), these phages were still detected by PCR whereas 434 435 they were inactivated. These results showed that UV treatment is highly efficient for the inactivation of *Microviridae* and that PCR methods using only a single target cannot be used 436 to evaluate this tertiary treatment. 437

Overall, similar reductions were observed for infectious Microviridae and SC, and the 438 proportion of infectious Microviridae was shown to be between 10% and 30% of total 439 infectious SC in both raw and treated WW. This suggested similar behaviour for un-tailed 440 Microviridae and tailed SC exposed to these WW treatments. A link between the phages' 441 442 morphology and the efficiency of WW treatment has been previously described for a full-scale MBR (Purnell et al., 2015). These authors only detected infectious Microviridae in treated WW 443 using transmission electronic microscopy; this was explained by tailed SC's higher adsorption 444 of solids and/or susceptibility to damage within the MBR process. The specific monitoring of 445 each *Microviridae* genus performed in our work suggested that  $\alpha$ 3-like genus might be more 446

sensitive to WW treatments than other *Microviridae*, but this point requires further
investigation taking account of the frequency of each genus.

449

450 **5. Conclusion** 

The PCR approach developed in the present study showed that the specific monitoring 451 452 of Microviridae genomes and infectious particles is possible in raw and treated WW; it could 453 also be applied to other environmental samples. In the WWTPs studied here, these phages appeared to behave similarly to all SC. Moreover, the size and capsid morphology of 454 Microviridae are closer to common pathogenic enteric viruses than other SC; this could be a 455 tremendous advantage for the evaluation of WW treatment. Thus, further studies focusing on 456 Microviridae may contribute to the improvement of WW treatment and reuse and, 457 458 consequently, human health risk management.

459

#### 460 **ACKNOWLEDGEMENTS**

This work was performed within the scope of Norostan, a project funded by the French Ministry of Agriculture and Food "CASDAR Recherche Technologique" (n°19ART374748). This study was also supported by the Actia VIROcontrol Joint Technological Unit.

The authors want to thank the "Syndicat Intercommunal du Bassin d'Arcachon" for kindly supplying raw and treated WW samples from WWTP. Additional support was obtained from Zone Atelier Moselle (ZAM).

467

#### 468 **COMPETING INTERESTS**

469 The authors declare no competing interest.

#### 471 **REFERENCES**

- 472 ASQAP, 2022. Australian Shellfish quality assurance program.
- 473 Aw, T.G., Howe, A., Rose, J.B., 2014. Metagenomic approaches for direct and cell culture
- evaluation of the virological quality of wastewater. J. Virol. Methods 210, 15–21.
- 475 https://doi.org/10.1016/j.jviromet.2014.09.017
- 476 Beck, S.E., Rodriguez, R.A., Hawkins, M.A., Hargy, T.M., Larason, T.C., Linden, K.G., 2016.
- 477 Comparison of UV-induced inactivation and RNA damage in MS2 phage across the
- 478 germicidal UV spectrum. Appl. Environ. Microbiol. 82, 1468–1474.
- 479 https://doi.org/10.1128/AEM.02773-15
- 480 Bellou, M., Kokkinos, P., Vantarakis, A., 2013. Shellfish-Borne Viral Outbreaks: A Systematic
- 481 Review. Food Environ. Virol. 5, 13–23. https://doi.org/10.1007/s12560-012-9097-6
- Bianchi, M.A.G., Bianchi, A.J.M., 1982. Statistical sampling of bacterial strains and its use in
  bacterial diversity measurement. Microb. Ecol. 8, 61–69.
- 484 https://doi.org/10.1007/BF02011462
- Blanch, A.R., Belanche-Muñoz, L., Bonjoch, X., Ebdon, J., Gantzer, C., Lucena, F., Ottoson, J.,
- 486 Kourtis, C., Iversen, A., Kühn, I., Mocé, L., Muniesa, M., Schwartzbrod, J., Skraber, S.,
- 487 Papageorgiou, G.T., Taylor, H., Wallis, J., Jofre, J., 2006. Integrated analysis of
- 488 established and novel microbial and chemical methods for microbial source tracking.
- 489 Appl. Environ. Microbiol. 72, 5915–5926. https://doi.org/10.1128/AEM.02453-05
- 490 Cantalupo, P.G., Calgua, B., Zhao, G., Hundesa, A., Wier, A.D., Katz, J.P., Grabe, M., Hendrix,
- 491 R.W., Girones, R., Wang, D., Pipas, J.M., 2011. Raw sewage harbors diverse viral
- 492 populations. MBio 2, 1–11. https://doi.org/10.1128/mBio.00180-11
- 493 Carducci, A., Battistini, R., Rovini, E., Verani, M., 2009. Viral removal by wastewater
- 494 treatment: Monitoring of indicators and pathogens. Food Environ. Virol. 1, 85–91.

495 https://doi.org/10.1007/s12560-009-9013-x

496	Chacón, L., Barrantes, K., Santamaría-Ulloa, C., Solano, M., Reyes, L., Taylor, L., Valiente, C.,
497	Symonds, E.M., Achí, R., 2020. A somatic coliphage threshold approach to improve the
498	management of activated sludge wastewater treatment plant effluents in resource-
499	limited regions. Appl. Environ. Microbiol. 86. https://doi.org/10.1128/AEM.00616-20
500	Cole, D., Long, S.C., Sobsey, M.D., 2003. Evaluation of F+ RNA and DNA Coliphages as Source-
501	Specific Indicators of Fecal Contamination in Surface Waters. Appl. Environ. Microbiol.
502	69, 6507–6514. https://doi.org/10.1128/AEM.69.11.6507-6514.2003
503	Dika, C., Duval, J.F.L., Francius, G., Perrin, A., Gantzer, C., 2015. Isoelectric point is an
504	inadequate descriptor of MS2, Phi X 174 and PRD1 phages adhesion on abiotic surfaces.
505	J. Colloid Interface Sci. 446, 327–334. https://doi.org/10.1016/j.jcis.2014.08.055
506	Environmental Protection Agency 40 CFR Parts 9, 141, and 142 National Primary Drinking
507	Water Regulations: Ground Water Rule; Final Rule rwilkins on PROD1PC63 with
508	RULES_2 (2006).
509	EPA Office of Water, Office of Science and Technology and Health and Ecological Criteria
510	Division (2015) Review of Coliphages as possible indicators of fecal contamination for
511	ambient water quality.
512	European comission (2020) Directive (EU) 2020/2184 of the european parliament and of the
513	council.
514	Farkas, K., Walker, D.I., Adriaenssens, E.M., McDonald, J.E., Hillary, L.S., Malham, S.K., Jones,
515	D.L., 2020. Viral indicators for tracking domestic wastewater contamination in the
516	aquatic environment. Water Res. https://doi.org/10.1016/j.watres.2020.115926
517	Fu, C.Y., Xie, X., Huang, J.J., Zhang, T., Wu, Q.Y., Chen, J.N., Hu, H.Y., 2010. Monitoring and
518	evaluation of removal of pathogens at municipal wastewater treatment plants. Water

- 519 Sci. Technol. 61, 1589–1599. https://doi.org/10.2166/wst.2010.757
- 520 Gantzer, C., Maul, A., Levi, Y., Schwartzbrod, L., 1998. Fate of the genome and infectious
- 521 units of coxsackie B3 virus in phosphate buffered saline. Water Res. 32, 1329–1333.
- 522 https://doi.org/10.1016/S0043-1354(97)00314-X
- 523 Gourmelon, M., Caprais, M.P., Ségura, R., Le Mennec, C., Lozach, S., Piriou, J.Y., Rincé, A.,
- 524 2007. Evaluation of two library-independent microbial source tracking methods to
- identify sources of fecal contamination in French estuaries. Appl. Environ. Microbiol. 73,
- 526 4857–4866. https://doi.org/10.1128/AEM.03003-06
- 527 Hartard, C., Rivet, R., Banas, S., Gantzer, C., 2015. Occurrence of and Sequence Variation
- among F-Specific RNA Bacteriophage Subgroups in Feces and Wastewater of Urban and
- 529 Animal Origins. Appl. Environ. Microbiol. 81, 6505–6515.
- 530 https://doi.org/10.1128/aem.01905-15
- Havelaar, A.H., Pot-Hogeboom, W.M., Furuse, K., Pot, R., Hormann, M.P., 1990. F-specific
- 532 RNA bacteriophages and sensitive host strains in faeces and wastewater of human and
- animal origin. J. Appl. Bacteriol. 69, 30–37. https://doi.org/10.1111/j.1365-
- 534 2672.1990.tb02908.x
- Health Canada (2019) Guidelines for Canadian Drinking Water Quality Enteric Viruses.
- 536 Ottawa, Ontario.
- 537 International Standardization Organization (2000) ISO-10705-2: Water Quality: Detection
- and Enumeration of Bacteriophages Part2: Enumeration of Somatic Coliphages.
- 539 Geneva,Switzerland.
- 540 ISO ISO 10705-2:2000 Water quality Detection and enumeration of bacteriophages —

541 Part 2: Enumeration of somatic coliphages (no date).

Jofre, J., Lucena, F., Blanch, A.R., 2021. Coliphages as a complementary tool to improve the

- 543 management of urban wastewater treatments and minimize health risks in receiving
- 544 waters. Water (Switzerland) 13. https://doi.org/10.3390/w13081110
- Jofre, J., Lucena, F., Blanch, A.R., Muniesa, M., 2016. Coliphages as model organisms in the
- 546 characterization and management of water resources. Water (Switzerland) 8, 1–21.
- 547 https://doi.org/10.3390/w8050199
- 548 Kelmer, G.A.R., Ramos, E.R., Dias, E.H.O., 2023. Coliphages as viral indicators in municipal
- 549 wastewater: A comparison between the ISO and the USEPA methods based on a
- systematic literature review. Water Res. 230, 119579.
- 551 https://doi.org/10.1016/j.watres.2023.119579
- 552 Korajkic, A., Kelleher, J., Shanks, O.C., Herrmann, M.P., McMinn, B.R., 2022. Effectiveness of
- 553 two wastewater disinfection strategies for the removal of fecal indicator bacteria,
- bacteriophage, and enteric viral pathogens concentrated using dead-end hollow fiber

ultrafiltration (D-HFUF). Sci. Total Environ. 831, 154861.

- 556 https://doi.org/10.1016/j.scitotenv.2022.154861
- 557 Korajkic, A., McMinn, B., Herrmann, M.P., Sivaganesan, M., Kelty, C.A., Clinton, P., Nash,
- 558 M.S., Shanks, O.C., 2020. Viral and bacterial fecal indicators in untreated wastewater
- across the contiguous united states exhibit geospatial trends. Appl. Environ. Microbiol.
- 560 86, 1–14. https://doi.org/10.1128/AEM.02967-19
- 561 Kotwal, G., Cannon, J.L., 2014. Environmental persistence and transfer of enteric viruses.
- 562 Curr. Opin. Virol. 4, 37–43. https://doi.org/10.1016/j.coviro.2013.12.003
- Levantesi, C., La Mantia, R., Masciopinto, C., Böckelmann, U., Ayuso-Gabella, M.N., Salgot,
- 564 M., Tandoi, V., Van Houtte, E., Wintgens, T., Grohmann, E., 2010. Quantification of
- 565 pathogenic microorganisms and microbial indicators in three wastewater reclamation
- and managed aquifer recharge facilities in Europe. Sci. Total Environ. 408, 4923–4930.

- 567 https://doi.org/10.1016/j.scitotenv.2010.07.042
- Lee, HS, 2009. Somatic coliphage families as potential indicators of enteric viruses in water
- and methods for their detection. PhD. University of North Carolina, Chapel Hill.
- Lucena, F., Duran, A.E., Morón, A., Calderón, E., Campos, C., Gantzer, C., Skraber, S., Jofre, J.,
- 571 2004. Reduction of bacterial indicators and bacteriophages infecting faecal bacteria in
- 572 primary and secondary wastewater treatments. J. Appl. Microbiol. 97, 1069–1076.
- 573 https://doi.org/10.1111/j.1365-2672.2004.02397.x
- 574 Martín-Díaz, J., Casas-Mangas, R., García-Aljaro, C., Blanch, A.R., Lucena, F., 2016. Somatic
- 575 coliphages as surrogates for enteroviruses in sludge hygienization treatments. Water
- 576 Sci. Technol. 73, 2182–2188. https://doi.org/10.2166/wst.2016.066
- Moreira, N.A., Bondelind, M., 2017. Safe drinking water and waterborne outbreaks. J. Water
   Health 15, 83–96. https://doi.org/10.2166/wh.2016.103
- 579 Muniesa, M., Lucena, F., Jofre, J., 1999. Study of the potential relationship between the
- 580 morphology of infectious somatic coliphages and their persistence in the environment.
- J. Appl. Microbiol. 87, 402–409. https://doi.org/10.1046/j.1365-2672.1999.00833.x
- 582 Muyzer, G., De Waal, E.C., Uitterlinden, A.G., 1993. Profiling of complex microbial
- 583 populations by denaturing gradient gel electrophoresis analysis of polymerase chain
- reaction-amplified genes coding for 16S rRNA. Appl. Environ. Microbiol. 59, 695–700.
- 585 https://doi.org/10.1128/aem.59.3.695-700.1993
- Nadkarni, M.A., Martin, F.E., Jacques, N.A., Hunter, N., 2002. Nadkarni MA 2002 Bacterial
- load by real time qPCR.pdf. Microbiology 148, 257–266.
- Nappier, S.P., Hong, T., Ichida, A., Goldstone, A., Eftim, S.E., 2019. Occurrence of coliphage in
- raw wastewater and in ambient water: A meta-analysis. Water Res. 153, 263–273.
- 590 https://doi.org/10.1016/j.watres.2018.12.058

- 591 Pecson, B.M., Ackermann, M., Kohn, T., 2011. Framework for using quantitative PCR as a
- 592 nonculture based method to estimate virus infectivity. Environ. Sci. Technol. 45, 2257–

593 2263. https://doi.org/10.1021/es103488e

- 594 Petrovich, M.L., Ben Maamar, S., Hartmann, E.M., Murphy, B.T., Poretsky, R.S., Wells, G.F.,
- 595 2019. Viral composition and context in metagenomes from biofilm and suspended
- 596 growth municipal wastewater treatment plants. Microb. Biotechnol. 12, 1324–1336.

597 https://doi.org/10.1111/1751-7915.13464

- 598 Purnell, S., Ebdon, J., Buck, A., Tupper, M., Taylor, H., 2015. Bacteriophage removal in a full-
- scale membrane bioreactor (MBR) Implications for wastewater reuse. Water Res. 73,
- 600 109–117. https://doi.org/10.1016/j.watres.2015.01.019
- Randazzo, W., Khezri, M., Ollivier, J., Le Guyader, F.S., Rodríguez-Díaz, J., Aznar, R., Sánchez,
- 602 G., 2018. Optimization of PMAxx pretreatment to distinguish between human norovirus
- with intact and altered capsids in shellfish and sewage samples. Int. J. Food Microbiol.

604 266, 1–7. https://doi.org/10.1016/j.ijfoodmicro.2017.11.011

- Reyes, A., Semenkovich, N.P., Whiteson, K., Rohwer, F., Gordon, J.I., 2012. Going viral: Next-
- generation sequencing applied to phage populations in the human gut. Nat. Rev.
- 607 Microbiol. 10, 607–617. https://doi.org/10.1038/nrmicro2853
- 608 Sidhu, J.P.S., Ahmed, W., Palmer, A., Smith, K., Hodgers, L., Toze, S., 2017. Optimization of
- sampling strategy to determine pathogen removal efficacy of activated sludge
- treatment plant. Environ. Sci. Pollut. Res. 24, 19001–19010.
- 611 https://doi.org/10.1007/s11356-017-9557-5
- Sidhu, J.P.S., Sena, K., Hodgers, L., Palmer, A., Toze, S., 2018. Comparative enteric viruses and
- coliphage removal during wastewater treatment processes in a sub-tropical
- environment. Sci. Total Environ. 616–617, 669–677.

- 615 https://doi.org/10.1016/j.scitotenv.2017.10.265
- 616 Simonet, J., Gantzer, C., 2006. Inactivation of poliovirus 1 and F-specific RNA phages and
- degradation of their genomes by UV irradiation at 254 nanometers. Appl. Environ.
- 618 Microbiol. 72, 7671–7677. https://doi.org/10.1128/AEM.01106-06
- 619 Sinclair, R.G., Jones, E.L., Gerba, C.P., 2009. Viruses in recreational water-borne disease
- outbreaks: A review. J. Appl. Microbiol. 107, 1769–1780.
- 621 https://doi.org/10.1111/j.1365-2672.2009.04367.x
- Truchado, P., Garre, A., Gil, M.I., Simón-Andreu, P.J., Sánchez, G., Allende, A., 2021.
- 623 Monitoring of human enteric virus and coliphages throughout water reuse system of
- wastewater treatment plants to irrigation endpoint of leafy greens. Sci. Total Environ.
- 625 782. https://doi.org/10.1016/j.scitotenv.2021.146837
- Turner, D., Kropinski, A.M., Adriaenssens, E.M., 2021. A roadmap for genome-based phage
   taxonomy. Viruses 13, 1–10. https://doi.org/10.3390/v13030506
- Young, S., Torrey, J., Bachmann, V., Kohn, T., 2020. Relationship Between Inactivation and
- 629 Genome Damage of Human Enteroviruses Upon Treatment by UV254, Free Chlorine,
- and Ozone. Food Environ. Virol. 12, 20–27. https://doi.org/10.1007/s12560-019-09411-
- 631 2
- 632 Wurtzer, S, Waldman, P, Ferrier-Rambert, A, Frenois-Veyrat, G, Mouchel, JM, Boni, M,
- 633 Maday, Y, OBEPINE consortium, Marechal, V, Moulin, L, 2021. Several forms of SARS-
- 634 CoV-2 RNA can be detected in wastewaters: Implication for wastewater-based
- epidemiology and risk assessment. Water Res. 198, 117183.
- 636 https://doi.org/10.1016/j.watres.2021.117183
- <sup>637</sup> Zhang, K., Farahbakhsh, K., 2007. Removal of native coliphages and coliform bacteria from
- 638 municipal wastewater by various wastewater treatment processes: Implications to

## 639 water reuse. Water Res. 41, 2816–2824. https://doi.org/10.1016/j.watres.2007.03.010



Figure 1: Quantification of viral infectious particles and genome copies in wastewater influents. (A) Localisation of the four WWTP in France. (B) Concentration values of infectious SC (grey pattern), infectious Microviridae (dark grey) and Microviridae genome copies (light grey) before WW treatments (bars represent the mean with SD for each condition). (C) Concentration values of infectious particles for the three genera of Microviridae in the influent (Box and whiskers representing the mean and the maximum and minimum values). (D) Concentration values of the genome of the three genera of Microviridae in the influent (Box and whiskers representing the mean and the maximum and minimum values). The LOQ of the three ddPCR assays in WW was between 4.7 log<sub>10</sub> GC/L. For all panels, Kruskal-Wallis statistical test was performed: stars represent significative differences (\* = p ≤ 0.05; \*\* = p ≤ 0.001 and \*\*\*\* = p ≤ 0.0001). In Figure 1B, the four WWTP were compared between each other and the WWTP were tested independently from each other for 1C and 1D.

Α



Figure 2: Variation in genome and infectious viral particles concentrations after WWTP treatments. (A)  $Log_{10}$  reduction of infectious SC (grey pattern), infectious *Microviridae* (dark grey) and *Microviridae* genome (light grey) after WW treatments (bars represent the mean with SD for each condition). (B)  $Log_{10}$  reduction of infectious *Microviridae* after WW treatments for each WWTP (bars represent the mean with SD for each condition). (C)  $Log_{10}$  reduction of the *Microviridae* genome copies after WW treatments for each WWTP (bars represent the mean with SD for each condition). (C)  $Log_{10}$  reduction of the *Microviridae* genome copies after WW treatments for each WWTP (bars represent the mean with SD for each condition). For all panels, Kruskal-Wallis statistical test was performed: stars represent significative differences (\* = p ≤ 0.05; \*\* = p ≤ 0.001; \*\*\* = p ≤ 0.001 and \*\*\*\* = p ≤ 0.0001). In Figure 2A, the WWTP were compared between each other and the WWTP were tested independently from each other for 2B and 2C.

### Table 1: Characteristics of the four WWTP

	WWTP-1	WWTP-2	WWTP-3	WWTP-4	
Location	Grand Est region	Grand Est region	Nouvelle Aquitaine region	Nouvelle Aquitaine region	
	(north-east of France)	(north-east of France)	(south-west of France)	(south-west of France)	
Capacity	400,000	440,000	150,000	135,000	
(inhabitant-					
equivalents)					
Type of WW	Urban	Urban	Urban	Urban	
	Residential	Residential and industrial <sup>a</sup>	Residential	Residential	
	Combined sewer system	Combined sewer-system	Separative sewer-system	Separative sewer-system	
Automated	Hach Lange Bühler	Endress Hauser	Endress Hauser	Endress Hauser	
sampler		(ASP2000)	(Liquistation CSF48)	(Liquistation CSF48)	
	Refrigerated	Refrigerated	Refrigerated	Refrigerated	
	Dry weather: 60 mL / 250-300 m <sup>3</sup>	60 mL / 200 m <sup>3</sup>	50 mL / 50 m <sup>3</sup>	50 mL / 50 m <sup>3</sup>	
	Wet weather: 60 mL / 350-450 m <sup>3</sup>				
Pre-treatment	Screening				
		Removal of sand, oil, and grease			
Treatment-I	Primary settling tank				
Treatment-II	Biological filtration				
	Secondary settling tank				
Treatment-III	Phosphorous treatment by flocculation with UV			IV	

<sup>a</sup>: industrial WW reached about 42,000 inhabitant-equivalents in terms of pollution.



Supplementary Figure 1: Concentration of the genome and infectious viral particles after WW treatment. (A) Concentration values of infectious SC (grey pattern), infectious *Microviridae* (dark grey) and *Microviridae* genome copies (light grey) after WWTP treatments (bars represent the mean with SD for each condition). (B) Concentration values of infectious particles for the genera of *Microviridae* in the effluents (Box and whiskers representing the mean and the maximum and minimum values)). (C) Concentration values of the genome of the three genera of *Microviridae* in the effluents (Box and whiskers representing the mean and the maximum and minimum values)). The LOQ of the three ddPCR assays in WW was ~4.7 log<sub>10</sub> GC/L. For all panels, Kruskal-Wallis statistical test was performed, stars represent significative differences (\* = p ≤ 0.05; \*\* = p ≤ 0.001; \*\*\* = p ≤ 0.001 and \*\*\*\* = p ≤ 0.0001). In Figure S1A, the four WWTP were compared between each other and the WWTP were tested independently from each other for S1B and S1C. For ddPCR, the LOQ was 5.0 log<sub>10</sub> gc/L